Accumulation of Iron, Manganese, Zinc and Cadmium by the Australian Freshwater Mussel Velesunio ambiguus (Phillipi) and Its Potential as a Biological Monitor

W. G. Jones^A and K. F. Walker^{B, C}

- ^A Centre for Environmental Studies, University of Adelaide, S.A. 5000, Present Address: 62 Channel Hwy., Taroona, Tas. 7006.
- ^B Department of Zoology, University of Adelaide, S.A. 5000.
- ^C To whom reprint requests should be addressed.

Abstract

The accumulation of iron, manganese, zinc and cadmium by freshwater mussels in the River Murray. South Australia, and their response to changes in environmental iron concentrations are considered. Metal loads varied markedly between individuals from the same population. The variability is accounted for partly by systematic relationships between metal loads and body weight and age, but not sex. The distribution of metals between the major organs is discussed, but the analysis of separate organs showed no advantage for biological monitoring. Comparisons between iron concentrations in river water and in mussels showed no clear correspondence. The study suggests that V, ambiguus may not be a good short-term monitor of iron, but still may have potential as a long-term and site-comparison monitor of metals, once inherent variability is taken into account.

Introduction

There is considerable interest in the use of organisms as biological monitors of pesticides and heavy metals (Wilhm 1975; Phillips 1977, 1978). One approach involves analysis of pollutant concentrations in organisms which act as bioaccumulators (Jones and Walker 1979). Although many organisms accumulate pollutants, only those having certain physical, ecological and physiological features have practical potential as monitors. A number of these do not involve the response of the organism to a specific pollutant, but are general characteristics, usually already known. Together they define the *prima facie* potential of the organism as a bioaccumulator monitor of any pollutant. Some of these features (size, abundance, hardiness, ease of sampling, mobility, tolerance to pollutants) have been discussed by Butler *et al.* (1971). In addition the organism should have a broad natural distribution, be taxonomically well defined, and its basic ecology should be known.

Velesunio ambiguus (Philippi) is an endemic Australian freshwater mussel (order Eulamellibranchiata, family Hyriidae) found throughout the Murray-Darling river system and in most coastal rivers of eastern and southern Australia, and with good prima facie potential as a bioaccumulator monitor (see Walker and Hillman 1977; Jones 1978). This paper assesses, in part, the practical potential of V. ambiguus as a monitor of iron, manganese, zinc and cadmium. The variability of metal concentrations among mussels in one population, distribution of metals between the major organs, and in situ response to changing concentrations of iron are considered.

Methods

Sample Collection and Preparation

Samples of 20–30 mussels were collected during March, April, August and September 1977 at Lock 3 on the River Murray, near Overland Corner, South Australia. They were taken from water 0.2–2 m deep and stored live out of water (for analysis within 3 weeks) or deep frozen.

The body was dissected from the shell using plastic forceps and a clean stainless steel scalpel. Wet and dry weights were measured to the nearest milligram, the latter after drying at 105 °C for 60 h. Homogenates of whole animals were made by grinding the dried body to a fine, consistent powder with a mortar and pestle. In the March sample separate organs were analysed whole, except the visceral mass which was homogenized in a tissue grinder.

The ages of European and North American freshwater mussels have been estimated by annual growth rings, mark-recovery experiments and length frequency plots (Haskin 1954; Coon *et al.* 1977). Although concentric rings are formed on the shells of V, ambiguax, these are not useful for aging (Walker and Jones, unpublished data), and mark-recovery experiments could not give useful results in one year. Length frequency plots could not be used because shell shape is variable. This method of ageing relies on animals of the same age growing at an even rate and being of similar shape so that length is an index of shell material produced. In V, ambiguax where shell shape varies between individuals, shell displacement volume is a better measure of shell material produced. Shell volume is, in fact, strongly correlated with shell weight (r = 0.98, n = 46), and so in this study was used as a qualitative measure of age.

Analysis of Tissue

A wet acid digestion, modified from Fowler and Oregioni (1976), was used. Homogenate subsamples of 0.5-0.8 g dry weight were digested in solutions of 2 mL concentrated HNO₃ and 0.3 mL concentrated HClO₄ for each 0.1 g dry weight. Digestion was at 80 °C for 1 h to minimize frothing and then at 150 °C until dry, fuming crystals remained. These were redissolved in 10 mL of a 1.1 mixture of 1.2 m HNO₃ and 1 m HCl, which in turn was diluted 10-fold with distilled deionized water for measurement of iron, manganese and zinc.

A Varian Techtron 1200 atomic absorption spectrophotometer was used for analysis by flame-absorption spectroscopy. Non-atomic absorption was significant only in cadmium measurement and was corrected for with a hydrogen continuum lamp. Mean recoveries from six spiked samples were (standard deviation in parentheses): iron 104°_{-0} (\pm 6), manganese 100°_{-0} (\pm 6), zinc 98°_{-0} (\pm 1), and cadmium 90°_{-0} (\pm 2). Results have not been corrected for recovery. Iron, manganese and zinc blanks were always less than 12°_{-0} of the sample concentrations, but early cadmium analyses were rendered unreliable by blanks of $30-100^{\circ}_{-0}$ typical sample concentrations. This was reduced to $5-102^{\circ}_{-0}$ by increasing the weight of tissue digested (to about 0.8 g dry weight) and washing digestion vessels in saturated chromic acid rather than 102°_{-0} HNO₃ as before. Two or three subsamples of each homogenate were analysed. Variability between subsamples was measured by the percentage coefficient of variation, calculated as $S/\overline{x} \times 100$, where S is the standard deviation and \overline{x} the mean. Average values (n=48) were iron 3.99°_{-0} , manganese 2.99°_{-0} , zinc 4.42°_{-0} , and cadmium (n=34) 10.59°_{-0} .

Analysis of Water

Water samples of 500 ml were collected, first twice weekly and then weekly, from the study site. They were preserved for metal analysis by immediately adding 3 ml of conc. HNO₃ (American Public Health Association 1975). Whole water samples were analysed by flame-absorption spectroscopy.

Results and Discussion

Variability of Metal Concentrations within a Population

Individuals of *V. ambiguus* from the same population accumulated varying concentrations of each of the metals studied (Table 1). High variability has been reported also in other studies of metal accumulation by molluscs (Bryan 1973: Ayling 1974; Phillips 1976a. 1976b; Bryan *et al.* 1977) and may hinder their use as monitors (Chow *et al.* 1976). However, some of this may be caused by systematic variation with factors such as body weight, age and sex (Phillips 1977).

ember 1977 at Lock 3 on water 0 · 2 2 m deep and

teel scalpel. Wet and dry or 60 h. Homogenates of h a mortar and pestle. In ich was homogenized in

nated by annual growth 1 ct al. 1977). Although ing (Walker and Jones, results in one year, hod of ageing relies on 1981 is an index of shell Il displacement volume plated with shell weight

ogenate subsamples of al concentrated HClO_ ien at 150 C until dry. 4 HNO3 and 1 M HCl. i, manganese and zinc. ir analysis by flameneasurement and was mples were (standard \pm cadmium 90% (\pm 2). ilways less than 1% of y blanks of 30-100% of tissue digested (to er than 10% HNO3 as ween subsamples was are S is the standard 9^{6} , zinc $4\cdot4^{6}$, and

study site. They were rican Public Health scopy.

nulated varying ability has been an 1973; Ayling use as monitors c variation with

Metal concentration was plotted against body weight on both linear and double logarithmic scales to show systematic variation. A linear correlation on the log-log scale indicates a power relationship between the variables (Boyden 1977). However, correlation coefficients (Table 2) show that linear equations describe the relationships

Table 1. Variability in metal concentration of V, ambiguus for three collection times N, number of mussels in sample; s.d., standard deviation; c.v., coefficient of variation $\binom{n}{n}$

				Concentration	ι (μg/g dry ·	w()
Sample	<i>N</i>		Iron	Manganese	Zinc	Cadmium
March	21	\overline{X}	5802	4222	311-5	
		s.d.	2934	2535	123.0	-
		c.v.	51	60	39	
August	23	\overline{X}	6764	4796	340 - 3	0.689
		s.đ.	2782	2580	126 - 7	0.631
		C.V.	41	53	37	92
Sept.	14	$.\overline{\chi^{i}}$	4589	4090	320 - 5	0.670
		s.d.	2625	2337	99 - 5	0.704
		C.V.	57	57	31	105

equally as well as power functions. For iron, manganese and zinc most relationships are statistically significant, although correlations often are not strong. Plots of metal concentration against shell volume also were made, and again linear equations are as good or better fits than power functions (Table 2). Although the relationships are not

Table 2. Correlation coefficients (r) for plots of V. ambiguus dry body weight against metal concentration and shell volume against metal concentration, on both linear and double logarithmic scales

* P = 0.05. ** P = 0.01. *** P = 0.001.

	Metal conen	v. body weight	Metal concn v. shell volum	
Metal	Linear	Log-log	Linear	Log-log
Iron				
Mar. $(n = 21)$	-0.46*	- 0 · 46*	0.46*	0.48*
Aug. $(n = 23)$	-0.56**	0 · 56**	0.28	0.31
Sept. $(n = 11)$	-0.52*	-0.46	0.30	0 - 32
Manganese				
Маг.	-0.46*	0 · 51**	0.42*	0.35
Aug.	0 - 38*	-0.40*	0.40*	0.29
Sept.	-0.43	-0.37	0.25	0.19
Zinc				
Mar.	- 0·75***	- 0 · 76***	-0.01	~ 0.01
Aug.	-0.33	-0.38*	0.39*	0.34
Sept.	0 - 54*	-0.54*	0.29	0.16
Cadmium				
Aug.	-0.42*	0·44*	-0.38*	() · 4()*
Sept.	() - ()9	0 · 23	-0.14	-0.17

strong, partial correlation coefficients showed that they were being masked by, and were in turn masking, the body weight-metal concentration relationships. Therefore multiple linear regressions of metal concentration against shell volume and body weight were calculated (Table 3).

Variation with body weight

Iron, manganese and zinc concentrations decrease systematically with increasing body weight, consistent with other studies of metals in molluses (Boyden 1977; Phillips 1977). Although the equations from month to month vary, the β weightings show that the effect of body weight remains fairly constant. The situation with cadmium is less clear. It seems that concentrations decrease systematically with body weight, but not as markedly as with the other metals.

Table 3. Summary of multiple regressions of *V. ambiguus* body weight and shell volume against metal concentration

The value of multiple R^2 multiplied by 100 estimates the percentage of the variation explained by the equation. The β weighting allows comparison of the relative effect of each independent variable on the dependent variable

W. dry weight (g): V, volume (cm³): * P = 0.05; ** P = 0.01; *** P = 0.001

Metal	Multiple R^2 (multiple R)	Parameters	Contribution to mult. R ²	eta weighting	Overall F
					O · CI · III
Iron					
Маг.	0.65	W'	0 · 44	-0.70	16 90***
n = 21	(0.81)	${\mathcal V}$	0.21	0.70	
Aug.	0 - 58	W	0.31	-0.76	13.80***
n = 23	(0.76)	ν	0 - 27	0.55	
Sept.	0.60	W	0 · 28	-0.79	6.09*
n = 11	(0.77)	\mathcal{V}	0.33	0.63	
Manganese					
Mar.	0 · 59	μ'	0.41	-0.68	12.71***
	$(0 \cdot 77)$	V	0-17	0.65	
Aug.	0.48	W	0-15	-0.60	9 · 3()***
-	(0.69)	V		0-61	
Sept.	0.42	W	0.19	- 0 - 66	2.84
·	(0.64)	V	0 · 23	0 - 53	
Zinc	, ,				
Mar.	0.63	W	0.63	- 0 · 85	15.44***
	(0.79)	V	0.00	0.29	
Aug.	0.40	W	0.11	0 - 53	6.66**
2	(0.63)	¥	0 - 29	0.58	0 00
Sept.	0.61	W	0 - 29	-0.80	6+30*
,	(0.78)	ν	0 · 32	0.63	0 30
Cadmium	, ,		- 		
Aug.	0:24	W	0 · 18	-0.33	3.15
<u>.</u>	(0.49)	ν	0.06	0 · 27	- 1.
Sept.	0.02	И∕	0.01	-0.05	0.09
	(0.15)	V	0.01	-0.12	0 02

Boyden (1974, 1977) investigated and reviewed the relationships between body weight and both metal content and concentration in bivalve molluscs, and concluded that these were best described by power functions. There are two obstacles to close comparison of the results of the present study with those of Boyden. First, the size range of mussels used here was 1·5–7 g dry weight. Boyden considered that a range of less than 10-fold usually gave non-significant relationships due to inherent biological variability. It is possible that use of a greater size range may reveal a power relationship as suggested by Boyden. Second, differences in age (shell volume) are masking the body weight-metal concentration relationship, necessitating multiple linear regressions. This treatment has not been necessary in other studies.

ally with increasing oyden 1977; Phillips eightings show that ith cadmium is less y weight, but not as

ll volume against metal

iation explained by the sendent variable on the

P = 0.001

70		
70 76 13·80*** 55 79 6·09* 63 68 12·71*** 65 50 9·30*** 51 56 2·84 53 65 15·44*** 19 13 6·66** 18 10 6·30* 3 3 3·15 7 5 0·09	ning	Overall F
70 76 13·80*** 55 79 6·09* 63 68 12·71*** 65 50 9·30*** 51 56 2·84 53 65 15·44*** 19 13 6·66** 18 10 6·30* 3 3 3·15 7 5 0·09		
76	70	16.90***
55 79 63 68 12·71*** 65 50 9·30*** 51 56 2·84 53 35 15·44*** 18 10 6·66** 18 10 6·30* 3 3 3·15 7 5 0·09	70	
79 6·09* 63 .68 12·71*** 65 50 9·30*** 51 56 2·84 53 .55 15·44*** 29 13 6·66** 18 10 6·30* 3 3 3·15 7 5 0·09		13.80***
63 .68 .12·71*** 65 50 .9·30*** 51 56 .2·84 53 .5 .15·44*** 29 .3 .6·66** .8 .0 .6·30* .3 .3 .3·15 .7 .5 .0·09		
12·71*** 65 50 9·30*** 51 56 2·84 53 15·44*** 19 13 6·66** 18 10 6·30* 3 3·15 7 5 0·09		6.09*
65 50 9·30*** 51 56 2·84 53 3 55 15·44*** 18 10 6·66** 18 10 6·30* 3 3 3·15 7 5 0·09	63	
65 50 9·30*** 51 56 2·84 53 3 55 15·44*** 18 10 6·66** 18 10 6·30* 3 3 3·15 7 5 0·09	68	19.71***
51 56 2·84 53 35 15·44*** 29 13 6·66** 18 10 6·30* 3 3 3·15 7 5 0·09		12 / 1
36	50	9.30***
53 55 15·44*** 29 13 6·66** 18 10 6·30* 3 3·15 7 5 0·09	51	
35		2.84
29 13 6 · 66** 18 10 6 · 30* 3 3 3 · 15 7 5 0 · 09	53	
29 13 6 · 66** 18 10 6 · 30* 3 3 3 · 15 7 5 0 · 09) C	1.7 4.40000
3 3·15 7 5 0·09	-	13:44***
3 3·15 7 5 0·09		6.66**
3 3·15 7 5 0·09	i8	0 00
3 3·15 7 5 0·09	:0	6.30*
5 0.09	3	
5 0.09		
5 0.09	.3 .2	3 · 15
2	/ E	0.00
))	0.09
	-	

bs between body s, and concluded bstacles to close n. First, the size d that a range of herent biological wer relationship hasking the body car regressions. Variation with age

With iron, manganese and zinc, the effect of age is opposite to that of body weight, as concentrations increase with increasing age. In contrast, cadmium concentrations decrease with increasing age, but as this trend contributes little to the multiple R^2 it cannot be considered significant. The effect of age on iron and manganese concentrations is fairly consistent, as indicated by the β weightings and contributions to the multiple R^2 . However, for zinc, the results from the March sample, which indicate that age has no significant effect upon concentrations, contrast with those for August and September. Only further study will explain this anomaly.

Unlike body weight, there has been little study of the effect of age on metal concentrations in molluscs. The results of Ayling (1974) showed that in Pacific oysters (Crassostrea gigas) the concentrations of cadmium and copper and probably also zinc and lead increase with age. Romeril (1971, cited in Raymont 1972) reported the same trend for zinc, copper and iron concentrations in the clam Mercenaria mercenaria, but later (Romeril 1974, cited in Boyden 1977) found that copper and iron concentrations in the same species decreased with age. Mackay et al. (1975) also claimed that concentrations of zinc, copper and cadmium in Crassostrea commercialis decreased with increasing age, but examination of their data suggests that this is not so.

The simplest explanation of metal concentration increasing with age is that metals are accumulated in excess of metabolic needs and excretory capacity, and that the excess is stored. As the animal gets older the amount of stored metal increases more rapidly than body weight.

There is evidence to suggest that trace metals in molluscs are partitioned into two pools: a mobile, readily exchangeable pool associated with low molecular weight compounds, and a permanent store tightly bound to proteins and/or in granules, removed from general metabolic circulation (Boyden 1977; Howard and Nickless 1977). Granules or granular cells containing metals have been found in a variety of molluscs (e.g. Harrison 1969; Bryan 1973; Coughtrey and Martin 1976; George et al. 1976; Bryan et al. 1977), and their function suggested as storage or excretion. Granules were isolated from V. ambiguus and investigated by Ch'ng-Tan (1968). Analysis showed the major metallic elements to be calcium (14.6%), manganese (5.4%) and iron (4.2%), while another seven metals, including zinc, were detected in trace quantities. She hypothesized that metal phosphates could account for up to 62% by weight of the granules and that they probably represented a store of metals that the animal was unable to excrete. Granules from a related freshwater mussel (Velesunio sp.) were analysed for zinc by Wood (1975) and were found to contain 211 μ g/g dry weight. Granules containing iron and manganese also have been reported in both European and North American freshwater mussels of the genus Anodonta (Dubuisson and Van Heuversuyn 1931, cited in Hobden 1970; Harrison 1969).

In molluscs the rates of excretion and metabolic turnover of both iron and manganese are slow (Harrison 1969; Hobden 1970). As the concentrations of iron, and to a lesser extent manganese and zinc, in the River Murray are high (Fig. 1), it is reasonable that *V. ambiguus* should accumulate them in excess, and store the excess in granules. This would explain why concentrations of these metals increase with age.

Variation with sex

Watling and Watling (1976) showed that different sexes of the marine mussel Choromytilus meridionalis accumulated different concentrations of several metals. For

comparison, separate regressions were calculated for male and female *V. ambiguus* in the August sample and compared by analysis of variance. There were no significant differences (Table 4), although the comparison of multiple regressions by analysis of variance is not a statistically sensitive technique.

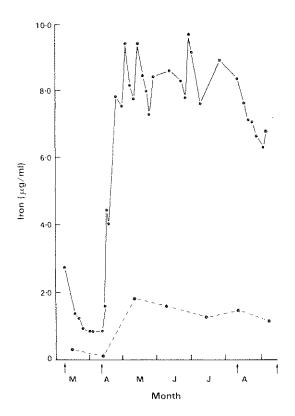


Fig. 1. Total iron concentrations in the River Murray at Lock 3, March to early September 1977. The broken line indicates data for 'available' iron (see text) in the Murray at Lock 9 (near the South Australian-Victorian border), supplied by the Engineering and Water Supply Department, South Australia. Arrows indicate when samples of mussels were collected.

Variation caused by gut contents

Some authors recommend starving bivalves before analysis to rid them of gut contents and pseudofaeces, as these may be sources of random variability (Hobden 1967; Karbe et al. 1975; George et al. 1976; Phillips 1976a; Bryan et al. 1977). V. ambiguus were kept alive out of water for at least 1 day after collection to allow pseudofaeces to accumulate at the edge of the mantle, where they could be removed easily during dissection. Variability in the visceral mass (see below) was more systematic than several other organs, and therefore gut contents probably were not an important source of random variation.

Metal Concentrations in Major Organs

The object here was to discover whether analyses of separate organs were advantageous for monitoring. Bryan (1973), for example, suggested that where whole body concentrations are difficult to detect, analysis of organs which accumulate higher concentrations than the body average may be useful. Another potential advantage, particularly relevant here, is that in some organs metal concentrations may show less random variation between individuals than concentrations for the whole body.

male *V. ambiguus* in were no significant ssions by analysis of

on concentrations urray at Lock 3, September 1977, e indicates data for (see text) in the 3 9 (near the South orian border). Engineering and Department, South ws indicate when sels were collected.

rid them of gut riability (Hobden n et al. 1977). V. ollection to allow could be removed clow) was more pably were not an

ate organs were that where whole ccumulate higher intial advantage, ns may show less whole body. Mussels from the March sample were used; organs analysed were the gills and labial palps, mantle, foot, kidney + heart, and the visceral mass, comprising the remaining organs and muscle.

Table 4. Comparison of multiple regressions for male and female V, ambiguus in the August sample Analyses of variance showed no differences between the concentrations accumulated by males and females. The range of body weights and volumes for each sex was similar, W, dry weight; V, shell volume

Metal	Male (n = 12)	Female (n = 11)	
Iron			
Average concn (µg,g) Standard deviation	6813 2870	6710 2822	
Multiple regression Manganese	Y = 2216 - 2015W + 261V	$Y = 12421 - 2850 \text{H}^{\circ} + 961^{\circ}$	
Average concn (µg g) Standard deviation Multiple regression	5153 3079 $Y = 1215 - 1798W + 2821$	4406 1975 Y = 7659 - 149911 + 451	
Zinc Average conen (µg,g) Standard deviation Multiple regression	$358 \cdot 5$ $162 \cdot 4$ $Y = 108 - 95 \cdot 0W + 13 \cdot 0V$	$320 \cdot 4$ $73 \cdot 8$ $Y = 421 - 36 \cdot 18^{\circ} + 0 \cdot 61^{\circ}$	
Cadmium Average concn (µg·g) Standard deviation Multiple regression	$0.722 \\ 0.551$ $Y = 2.18 - 0.18W - 0.02F$	$7 = 421 + 36 \cdot 13 + 0 \cdot 63$ $0 \cdot 652$ $0 \cdot 736$ $Y = 3 \cdot 01 + 0 \cdot 3537 + 0 \cdot 0237$	

As with metal concentrations in the whole body there is considerable variation between individuals (Table 5). Therefore the concentrations in individual organs were

Table 5. Average metal concentrations in the major organs of V, ambiguus March sample, n = 21, s.d., standard deviation; c.v., coefficient of variation $\binom{9}{5}$

Organ	Concentration $(\mu g/g)$ Iron Manganese Zinc			
Gills + palps	10417	17950	915	
s.d.	3374	8140	321	
C.V.	32	45	3.5	
Kidney + heart	7737	8405	822	
s.d.	8514	5249	1619	
C.V.	105	62	196	
Foot	2918	535	221	
s.d.	2600	362	72	
C.V.	89	68	33	
Mouth	10206	6488	383	
s.đ.	5985	4782	196	
c.v.	59	74	51	
Visceral mass	4572	1606	192	
s.d.	2760	943	69	
C.V.	60	59	36	

plotted against whole body weight and shell volume to show systematic variation (Table 6). Variation in concentrations in the mantle was the most systematic and the most likely to warrant further investigation. However, multiple correlation was not as

high as for whole body concentrations (Jones, unpublished data). Hence from the point of view of reducing random variability there is no virtue in analysing organs separately.

Table 6. Correlation coefficients (r) for relationships of metal concentrations in organs with whole body dry weight and shell volume

The correlations between whole body metal concentrations and these variables are shown for comparison. March sample, n = 21; * P = 0.05; ** P = 0.01; *** P = 0.001

		Whole body dry weight v.		Shell volume v.		
Metai	Organ	Metal concn in organs	Metal concn in whole body	Metal concu in organs	Metal conen in whole body	
Iron	Gills + palps	0 · 08		0 · 31		
	Kidney + heart	0.22		0.58**		
	Foot	— () · 37*	-0.46*	0.08	0.46*	
	Mantle	0 · 53**		0.42*		
	Visceral mass	() · 4()*		0.40*		
Manganese	Gills + palps	0.00		0.40*		
	Kidney + heart	-0.04		0.57**		
	Foot	-0.40*	-0.45*	-0.02	0.42*	
	Mantle	-0.50**		0.23		
	Visceral mass	-0.52**		0.28		
Zinc	Gills + palps	-0.12		0.03		
	Kidney + heart	-0.32		0.05		
	Foot	-0.37*	-0.75***	-0.26	0.00	
	Mantle	-0.79***		-0.10		
	Visceral mass	-0.67***		-0.19		

The patterns of iron, manganese and zinc distribution in V, ambiguus (Tables 5 and 7) are broadly similar to those reported in other freshwater mussels. Iron concentrations in the gills and mantle, relative to the rest of the body, are higher than those in the North American species *Elliptio complanata* (Hobden 1970). Also the

Table 7. Average percentage of total body metal content in major organs of *V. ambiguus*

March sample. n = 21. Standard deviations in parentheses

Organ	Iron	Manganese	Zinc	
Gills + palps	20 - 3	48 · 7	32.5	
	(4:3)	(11.3)	(9.8)	
Kidney + heart	3.0	5 · 1	5-3	
	(2-1)	(3.2)	(5.9)	
Foot	3 · 8	1.0	3 · 7	
	(5.2)	(1.1)	(1.9)	
Mantle	20 · 7	18.5	14 · 7	
	(6.1)	(7.4)	(3.8)	
Visceral mass	52 - 2	26 · 7	43 · 8	
	(10.2)	(8.8)	(11-5)	

average concentration in *V. ambiguus* is two orders of magnitude higher, reflecting high iron concentrations in the River Murray. For manganese the distribution pattern is almost identical with that recorded in other species, with the dominant feature being very high concentrations in the gills (Harrison 1969; Seah and Hobden 1969). The

a). Hence from the in analysing organs

tans with whole body dry

shown for comparison.

001

hell volume F.
ncn Metal conen
ns in whole body

**

0-46*

*

0-42*

guus (Tables 5 and er mussels. Iron ly, are higher than n 1970). Also the

organs of

Zinc
32+5
(9-8)
5 - 3
(5-9)
3 · 7
$(1 \cdot 9)$
4-7
[3:8)
·3 · 8
1 · 5)

higher, reflecting stribution pattern ant feature being obden 1969). The calcareous tissue in the gills of *Anodonta nuttaliana*, found by Harrison to contain extremely high concentrations of manganese, was not evident in *V. ambiguus* and for sake of comparison was considered part of the gills. The only notable feature of the distribution of zinc is that the concentration in the kidney+heart is much higher, relative to the other organs, than in other species (Harrison 1966, 1969; Pauley and Nakatani 1968).

The gills, and to a lesser extent the mantle, have high concentrations of all metals. They also contain large patches of granules. The foot, which contains no granules (Ch'ng-Tan 1968), has the lowest metal concentrations of all the organs analysed. These facts are consistent with the distribution of metals amongst the major organs being largely associated with the distribution of storage granules.

Response of Mussels to Changing Concentrations of Iron in situ

Water samples were collected once or twice weekly between March and September 1977, and analysed for total iron concentrations so that the average concentrations in the monthly samples of mussels could be compared with any changes in river concentrations (Fig. 1). Data on changes in water temperature, suspended and dissolved solids and flow rate also were obtained. Concentrations of biologically available iron (soluble or loosely bound and leached by acetic acid) measured at Lock 9 on the River Murray were supplied by the South Australian Engineering and Water Supply Department.

Differences in the average metal concentrations of mussel samples are not readily shown by the comparison of multiple regressions. To show this, values for iron for a 'standard' mussel of $3 \cdot 5$ g dry weight and 40 cm^3 shell volume (about the middle of the size range for all samples) were derived from multiple regressions of iron concentration against body weight and shell volume. Sample sizes were n = 21 (March), n = 14 (April), n = 23 (August), and n = 11 (September), and iron concentrations (in $\mu g/g$) were 5574, 4746, 5199 and 5211 for these months respectively.

Although concentrations in the standard mussel rose slightly from April to August and September, none of the monthly regressions were significantly different. The sharp increase in both total and available iron during mid-April and sustained throughout the sampling period, was not reflected by concentrations in *V. ambiguus*. This is not readily explained. Although the activity of mussels may be reduced by cold temperatures, animals were found filtering during May, August and September, and it seems unlikely that cooler water temperatures (minimum 10.5 °C in July) can fully explain the lack of response. Alternatively, the mussels may closely regulate their iron content, but this seems incongruous with the high concentrations that they contain and the large number of storage granules. It is possible that an increase in the concentrations in the mussels occurred, but was small in comparison to the base levels already present and was not detectable above the background variability. This seems most likely, perhaps in conjunction with other factors.

Conclusions

The variability of metal concentrations between individuals of *V. ambiguus* must limit its usefulness as a monitor. Some systematic variation can be taken into account by careful choice of samples within limited size ranges and/or normalization using regressions (e.g. Phillips 1976a), but a large unpredictable component remains. This is not obviated by the analysis of any of the major organs or by separating sexes, and is

unlikely to be markedly reduced by starving animals before analysis. The amount of variability is not unusual, being similar to that in most other molluses analysed for heavy metals (e.g. Chow et al. 1976; Phillips 1976a, 1976b; Coughtrey and Martin 1977), Apart from reducing the sensitivity of comparisons, it necessitates large sample sizes which increase the time and expense of analyses. In routine monitoring this disadvantage may be lessened by pooling a large number of animals and analysing homogenate subsamples (Mackay et al. 1975).

The correlation of iron, manganese and perhaps zinc concentrations with age has been attributed to a permanent store of excess metal. When permanently stored metal accounts for a large proportion of the total metal concentration, it may reduce the apparent responsiveness of mussels to changes in ambient concentrations. This in turn increases the likelihood that the response will be obscured by variability. Therefore the potential of *V. amhiguus* as a short-term monitor may be further limited for metals which are accumulated in a large permanent store. On the other hand, it may be useful as a long-term and site comparison monitor of such metals. Permanently stored metal may be avoided by analysis of the foot, in which metal concentrations have no correlation with age.

Finally, the lack of response of V, ambiguus to large increases in the ambient concentrations of biologically available iron means that it would not be a good short-term monitor of this metal. Controlled laboratory experiments of metal uptake and excretion are needed to further investigate its response to other metals, and to explain more clearly the response to iron.

Acknowledgment

We thank Professor W. D. Williams for allowing one of us (W. G. J.) the use of space and facilities in the Zoology Department. University of Adelaide. The work was undertaken as part of a M.Env.St. thesis by W. G. J., and financial support was provided through the Centre for Environmental Studies, University of Adelaide.

References

- American Public Health Association (1975). 'Standard Methods for the Examination of Water and Wastewater.' 14th Edn. (A.P.H.A.: Washington.)
- Ayling, G. M. (1974). Uptake of cadmium, zinc, copper, lead and chromium in the Pacific oyster. Crassostrea gigas grown in the Tamar River, Tasmania. Water Res. 8, 729–38.
- Boyden, C. R. (1974). Trace element content and body size of molluses. *Nature*, (London). 251, 311-14. Boyden, C. R. (1977). Effect of size upon metal content of shellfish. *J. Mar. Biol. Assoc. U.K.* 57, 675-714.
- Bryan, G. W. (1973). The occurrence and seasonal variation of trace metals in scallops, *Pecten maximus* (L) and *Chlamys opercularis* (L). *J. Mar. Biol. Assoc. U.K.* 53, 145-66.
- Bryan, G. W., Potts, G. W., and Forster, G. R. (1977). Heavy metals in the gastropod mollusc Huliotis tuberculata (L). J. Mar. Biol. Assoc. U.K. 57, 379-90.
- Butler, P. A., Andren, L., Bonde, G. J., Jerenelov, A., and Reisch, D. J. (1971). Monitoring organisms. In 'FAO Technical Conference on Marine Pollution and its effects on living resources and fishing. Rome. 1970. Supplement 1: Methods of detection, measurement and monitoring of pollutants in the marine environment'. (Ed. M. Ruvio.) pp. 101–112. (Fishing News (Books) Ltd: London.)
- Ching-Tan, K. S. (1968). Some aspects of renal and excretory physiology in the Australian freshwater mussel. *Velesunio ambiguus* (Philippi), (Mollusca, Bivalvia). Ph.D. Thesis, Monash University.
- Chow, T. J., Snyder, H. G., and Snyder, C. B. (1976). Mussels (*Mytilus* sp.) as an indicator of lead pollution. *Sci. Total Environ.* 6, 53-63.
- Coon, T. G., Eckblad, J. W., and Trygstad, P. M. (1977). Relative abundance, and growth of mussels (Mollusca; Eulamellibranchia) in pools 8, 9 and 10 of the Mississippi River. *Freshwater Biol.* 7, 279-85.
- Coughtrey, P. J., and Martin, M. H. (1976). The distribution of Pb. Zn, Cd. and Co within the pulmonate molluse Helix aspersa, Muller. Oecologia 23, 315-22.

sis. The amount of fluses analysed for antrey and Martin itates large sample to monitoring this hals and analysing

tions with age has aently stored metal it may reduce the ttions. This in turn ility. Therefore the limited for metals id, it may be useful iently stored metal atrations have no

es in the ambient of be a good shortmetal uptake and als, and to explain

G. J.) the use of ide. The work was acial support was ty of Adelaide.

nation of Water and

in the Pacific oyster.

London). **251**, 311–14. soc. U.K. **57**, 675–714. ss. Pecten maximus (L)

opod molluse Haliotis

nitoring organisms. In res and fishing, Rome, flutants in the marine on.)

Australian freshwater ush University. attor of lead pollution.

nd growth of mussels matter Biol. 7, 279-85, within the pulmonate

Coughtrey, P. J., and Martin, M. H. (1977). The uptake of lead, zinc, cadmium and copper by the pulmonate molluse, *Helix aspersa* Muller, and its relevance to the monitoring of heavy metal contamination of the environment. *Oecologia* 27, 65–74.

Fowler, S. W., and Oregioni, B. (1976). Trace metals in mussels from the N.W. Mediterranean. *Mar. Pollut. Bull.* 7, 26-9.

George, S. G., Pirie, B. J. S., and Coombes, T. L. (1976). The kinetics of accumulation and excretion of ferric hydroxide in *Mytilus edulis* (L.) and its distribution in the tissues. *J. Exp. Mar. Biol. Ecol.* **23**, 71–84. Harrison, F. L. (1966). Metabolism of Mn⁸⁴ and other eations in the freshwater clam. *The Physiologist* **9**, 200 (Abstract only).

Harrison, F. L. (1969). Accumulation and distribution of ⁸⁴Mn and ⁶⁸Zn in freshwater clams. Proc. 2nd Nat.
 Sympos. Radioecology, pp. 198–220.

Haskin, H. H. (1954). Age determination in molluses. New York Acad. Sci. Ser. 2, 16, 300-4.

Hobden, D. J. (1967). Iron metabolism in Mytilus edulis. 1. Variation in total content and distribution. J. Mar. Biol. Assoc. U.K. 47, 597-606.

Hobden, D. J. (1970). Aspects of iron metabolism in a freshwater mussel. Can. J. Zool. 48, 83-6.

Howard, A. G., and Nickless, G. (1977). Heavy metal complexation in polluted molluses. II. Oysters (Ostrea edulis and Crassostrea gigas). Chem-Biol. Interact. 17, 257–64.

Jones, W. G. (1978). The freshwater mussel, Velesunio ambiguus (Phillippi), as a biological monitor of heavy metals. Master of Environmental Studies Thesis, University of Adelaide.

Jones, W. G., and Walker, K. F. (1979). An outline of biological monitoring in aquatic environments. Water 6(2), 17-19.

Karbe, L., Antonacopoulos, N., and Schnier, C. (1975). The influence of water quality on the accumulation of heavy metals in aquatic organisms. Verh. Int. Verein Limnol. 19, 2094–101.

Mackay, N. J., Williams, R. G., Kacprzac, J. L., Kazacos, M. N., Collins, A. J. and Auty, E. H. (1975).
 Heavy metals in cultivated oysters (Crassostrea commercialis — Saccostrea cucullata) from the estuaries of New South Wales. Aust. J. Mar. Freshwater Res. 26, 31–46.

Pauley, G. B., and Nakatani, R. E. (1968). Metabolism of the radioisotope 65Zn in the freshwater mussel Anodonta californiensis. J. Fish. Res. Board Can. 25, 2691–4.

Phillips, D. J. H. (1976a). The common mussel, Mytihis edulis as an indicator of pollution by zinc, cadmium, lead and copper. I. Effects of environmental variables on uptake of metals. Mar. Biol. 38, 59-69.

Phillips, D. J. H. (1976b). The common mussel, *Mytihus edulis* as an indicator of pollution by zinc, cadmium, lead and copper. II. Relationship of metals in the mussel to those discharged by industry. *Mar. Biol.* 38, 71–81.

Phillips, D. J. H. (1977). The use of biological indicator organisms to monitor trace metal pollution in marine and estuarine environments—a review. *Environ. Pollut.* 13, 281–317.

Phillips, D. J. H. (1978). Use of biological indicator organisms to quantitate organochlorine pollutants in aquatic environments—a review. *Environ. Pollut.* 16, 167–229.

Raymont, J. E. G. (1972). Some aspects of pollution in Southampton Water, *Proc. R. Soc. London.* Ser. B. 180, 451–68.

Seah, T. C. M., and Hobden, D. J. (1969). Manganese in the freshwater clam. Can. J. Biochem. 47, 557-60.
 Walker, K. F., and Hillman, T. J. (1977). Limnological survey of the River Murray in relation to Albury-Wodonga, 1973-1976. (Gutteridge, Haskins and Davey and Albury-Wodonga Development Corporation: Albury.)

Watling, H. R., and Watling, R. (1976). Trace metals in *Choromytilus meridionalis*. Mar. Pollut. Bull. 7, 91-5. Wilhm, J. L. (1975). Biological indicators of pollution. In 'River Ecology'. (Ed. B. A. Whitton.) pp. 375–402. (Blackwell: Oxford.)

Wood, K. F. (1975). Some aspects of the accumulation of zinc by an Australian species of freshwater mussel (subfamily Velesunioniae). (*Velesunio* species). B. Ag. Sci. Thesis, LaTrobe University.

Manuscript received 9 April 1979, accepted 23 July 1979